

Removal of pharmaceutical compounds by electrochemical processes in real wastewater

Ensano B.M.B.¹, Borea L.², Naddeo V.^{2*}, Belgiorno V.², De Luna M.D.G.^{1,3}, Ballesteros, F.C. Jr.^{1,3}

¹Environmental Engineering Program, National Graduate School of Engineering, University of the Philippines, 1101 Diliman, Quezon City, Philippines

²Sanitary and Environmental Engineering Division (SEED), Department of Civil Engineering, University of Salerno, Fisciano 84084 (SA), Italy

³Department of Chemical Engineering, University of the Philippines, 1101 Diliman, Quezon City, Philippines

*corresponding author: Benny Marie Ensano

e-mail: bmensano@yahoo.com

Abstract.

This study investigated the viability of treating pharmaceuticals from real municipal wastewater using process electrochemical i.e. intermittent electrocoagulation. Batch experiments were performed using a laboratory scale electrochemical reactor with aluminum and stainless steel as anode and cathode, respectively. Diclofenac (DCF), carbamazepine (CBZ) and amoxicillin (AMX) were selected as representative of highly consumed pharmaceuticals under different drug categories (anti-inflammatory, antiepileptic and antibiotic) that are frequently detected in water resources. The effects of continuous and intermittent (5 min ON/ 20 min OFF) application of direct current (DC) field on the abatement of pharmaceutical compounds were determined and found to favor the latter in terms of treatment efficiency at varying current densities (CD = 0.3, 0.5 and 1.15 mA/cm^2). The removal efficiencies of conventional pollutants namely chemical oxygen demand (COD), dissolved organic carbon (DOC), aromatic and humic substances (in terms of UV₂₅₄), ammonia nitrogen (NH₄-N) and orthophosphate (PO₄-P) were also obtained and found to be directly proportional to current density and reaction time.

Keywords: diclofenac (DCF); carbamazepine (CBZ); amoxicillin (AMX); emerging contaminants (ECs); electrocoagulation

1. Introduction

Pharmaceutical compounds are emerging contaminants whose occurrence in the aquatic ecosystems has elicited serious concerns due to their adverse chronic toxicological effects on living organisms including human beings (Houtman *et al.*, 2014). Since conventional biological treatment methods remain inadequate to completely degrade recalcitrant pharmaceuticals from contaminated wastewater (Inyang *et al.*, 2015), advanced wastewater treatment technologies should therefore be explored. It has been demonstrated that electrochemical processes offer an attractive approach for treating refractory pollutants from wastewater due to their distinct advantages such as simple, highly effective and lesser footprint (Ensano *et al.*, 2017).

In addition, the significant improvement of the electrode materials and their coupling with low-cost renewable energy resources make electrochemical processes a more appealing method. In an electrochemical system using a sacrificial anode, electrocoagulation is the main mechanism of pollutant removal (Ensano et al., 2016). The electrochemical dissolution of the anode generates various monomeric and polymeric metal ion complexes which react with contaminants in the wastewater forming flocs that can both destabilize and aggregate colloidal particles or precipitate and adsorb dissolved compounds (Elabbas et al., 2016; Ganzenko et al., 2014). In the case of aluminum anode, Al³⁺ ions, formed from the electro-oxidation of the anode, react with OH⁻ ions, generated from water reduction at the cathode, producing amorphous Al(OH)₃(s) flocs. The application of electricity in the system also generates bubbles due to the evolution of oxygen gas and hydrogen gas at the anode and cathode, respectively. These bubbles contribute to the collection of generated flocs by electroflotation process (Liu et al., 2015). Heavier flocs settle at the bottom thru sedimentation. Electrocoagulation has been used as cheap alternative to chemical coagulation for wastewater treatment. Aside from its low operating costs, electrocoagulation has compact treatment facility, faster rate of pollutant removal and lower production of sludge (Kobya Demirbas 2015). and Additionally. electrocoagulation has better treatment efficiency than chemical coagulation since the latter produces unwanted anions, such as chloride and sulfide ions, that may compete with cations in the solution and therefore affect the removal of contaminants (Pérez et al., 2015). Flocs produced from electrocoagulation are more stable and contain less bound water which can be removed from the aqueous medium either by sedimentation, flotation or filtration (Elabbas et al., 2016). Electrocoagulation has been successfully applied to various types of wastewater. In this study, the performance of electrochemical system in the elimination of pharmaceuticals from real municipal wastewater was examined. DCF, CBZ, and AMX were selected as representatives of therapeutic groups most frequently found in waters. The effect of varying experimental conditions such as initial concentration, reaction time and current density on the removal of these

compounds was studied. Intermittent application of minute electric field was also employed to reduce energy consumption and operational cost. To the best of the authors knowledge, this is the first paper which applies intermittent electrocoagulation on the removal of recalcitrant pharmaceuticals from raw wastewater.

2. Materials and Methods

2.1. Chemicals and Reagents

Analytical grade (purity > 99 %) DCF ($C_{14}H_{10}Cl_2NNaO_2$), CBZ ($C_{15}H_{12}N_2O$) and AMX ($C_{16}H_{19}N_3O_5S \cdot 3H_2O$) were manufactured by Sigma-Aldrich and were used in this study without any further purification. A specified amount (0.01 mg/L) of each pharmaceutical was spiked into 1.8 L raw wastewater that was obtained from the effluent of primary clarifier in the municipal wastewater treatment plant of Salerno, Italy. The feed wastewater has the following composition: COD (1.7 mg/L), UV₂₅₄ (0.207), NH₄-N (286.55 mg/L), NO₃-N (3.27 mg/L), PO₄-P (2.18 mg/L), pH (7.4), T (19 °C), redox potential (205) and conductivity (1780 μ S/cm).

2.2. Experimental Set-up

All the experiments were carried out in a 2 L electrochemical reactor that contains aluminum anode (5 cm x 16.5 cm x 0.2 cm) and stainless steel cathode (6 cm x 16.5 cm x 0.2 cm) set 5 cm apart. A direct current was supplied by a DC regulated power source (TTi CPX400S 420W DC Power Supply). A mixing speed of 300 rpm was maintained using magnetic stirrer to ensure uniform concentration within the reactor. Before the start of each experiment, the electrodes were cleaned using sodium hypochlorite and was air-dried.

2.3. Laboratory Analyses

Dissolved oxygen concentration (DO), pH, temperature, conductivity and redox potential were measured using a multiparametric probe (Hanna Instruments, HI769828) before, during and after each run to understand the behavior of the process. Samples were collected from the influent and effluent and were analysed in terms of COD, DOC, UV254, NH4-N and PO4-P according to standard methods (APAT, 2003). The amount of pharmaceutical compounds present in the samples were then analyzed using 4000Q Trap LC-MS/MS System (Applied Biosystems, Foster City, USA) in ESI-positive mode with a mobile phase composed of A: 0.1% formic acid in water and B: acetonitrile-water (1:1, v/v) solution (limit of quantification lower of 1 ng/L). The method detection limit (MDL) was between 0.9 and 8 ng/L in spiked water samples. The precision of the method, calculated as relative standard deviation, ranged from 0.9 to 3.0%.

3. Results and Discussion

3.1 Removal of conventional pollutants

The removal efficiencies of conventional pollutants (COD, DOC, UV_{254} , NH₄-N and PO₄-P) are reported in Table 1. A total of 9 runs were performed at varying current densities (0.3, 0.5 and 1.15 mA/cm²) and HRT (3, 6 and 19h). From

the results observed, it can be generally construed that the treatment of the conventional pollutants is directly proportional to that of current density and HRT. The largest removal was obtained using 1.15 mA/cm² current density and 19 h reaction time. The removal of organic pollutants namely COD and DOC are largely attributed to the action of coagulants in the system. Aluminum metal complexes, that are formed from the dissolution of the sacrifical anode, reacts with the contaminants in the wastewater forming flocs which are helpful for the fast adsorption of dissolved organic compounds (Tafti et al., 2015). UV_{254} is a parameter used to estimate the dissolved aromatic fraction of natural organic matter (NOM) as well as the humic substances in water samples (Weishaar et al., 2003). During electrocoagulation, high molecular weight fraction of NOM is degraded into low molecular weight and more hydrophilic compounds (Mohora et al. 2014). Also, the functional groups of humic acid are attracted to positive Al species which facilitates their removal (Ulu et al. 2015). The decrease of NH₄-N concentration after the electrochemical treatment is due to the oxidation of ammonia molecules at the anode to nitrate (Lin and Wu 1996). Moreover, 100% of PO₄-P concentration was removed regardless of the current density applied. This can be ascribed to the adsorption of the soluble phosphorus in the mixed liquor by the generated Al coagulants and the precipitation of phosphate ions into AlPO_{4(s)} and Al₆(OH₁₅)PO_{4(s)} (Ensano et al., 2016)

3.2 Removal of pharmaceutical compounds

Figure 1 shows the effects of continuous and intermittent application of electricity on the removal of selected pharmaceuticals from municipal wastewater. Using 0.01 mg/L initial concentration of each compound and HRT of 3 hours, it can be seen from the figure that, at lower current density (0.3 mA/cm^2) , the system with continuously applied DC field obtained higher abatement of the recalcitrant micropollutants than the intermittent system. However, as current density increases, the removal of the compounds favors the intermittent system. While both the continuous and intermittent applications generally obtained an increasing trend as CD increases, the difference between their treatment efficiencies also improves favoring the latter system. As discussed previously, the mechanisms involved during conventional electrocoagulation include the dissolution of the sacrificial anode, the formation of coagulants and the removal of pollutants by adsorption onto Al(OH)3 flocs, charge neutralization and electroflotation. These processes occur simultaneously in a system with continuously applied electric field. Nevertheless, in an intermittent application, these processes are initiated when the power is ON (5 min) and then the power gap (20 min OFF) allows further removal of the micropollutants hence maximizing the adsorption and neutralization potential of generated coagulants. As electricity intensifies, the amount of coagulants also increases, therefore, the higher removal. These results are in accordance with previously published literatures (Ensano et al., 2017, Borea et al., 2017). To better understand which removal mechanism works better in each pharmaceutical compound, the knowledge on their

Table 1. Removal of conventional pollutants at varying current density and HRT.

Current Density (mA/cm ²)	HRT (h)	% Removal				
		COD	DOC	UV ₂₅₄	NH ₄ -N	PO ₄ -P
0.3	3	17%	7%	28%	18%	100%
	6	9%	18%	45%	12%	100%
	19	14%	35%	42%	26%	100%
0.5	3	17%	11%	31%	21%	100%
	6	13%	23%	40%	25%	100%
	19	27%	31%	50%	19%	100%
1.15	3	9%	10%	44%	12%	100%
	6	14%	21%	59%	19%	100%
	19	25%	30%	73%	17%	100%

physicochemical properties is important. The low hydrophobicity of the selected pharmaceuticals ($K_{ow DCF} =$ 0.70, $K_{\rm ow\,AMX}$ = 0.87, $K_{\rm ow\,CBZ}$ = 2.45) suggests that they are more likely to be removed via charge neutalization rather than the adsorption on coagulants (Phan et al., 2015). In this study, the pH of the solution (pH_{ave}=7.4) was not adjusted in order to simulate the typical pH of municipal wastewater. Since the acid dissociation constant (pKa) of DCF is 4.2 (Acero et al., 2016), it is anionic at neutral pH. The uptake of this compound via electrocoagulation is attributed to the electrostatic attraction between DCF and the positive aluminum complexes. CBZ has two pKas: $pKa_1 = 2.3$ is related to the protonation of the NH₂ group; and pKa2=13.9 is related to the deprotonation of the amino group (Rao et al., 2016). At neutral pH, CBZ is in the molecular state and shows hydrophobic property (Nghiem et al., 2006). Hence, among the three pharmaceuticals, CBZ has the highest tendency to be adsorbed on the Aluminum coagulants rather than charge neutralization. AMX, on the other hand, has three pKa values namely pKa₁=2.68, pKa₂=7.49 and pKa₃=9.63 which are attributed to the ionization of its carboxyl, amine and phenolic hydroxyl groups, respectively (Hu et al., 2016). At pH~7, the caboxyl groups are transformed into COO⁻ while amine and phenol groups remains neutral (Putra et al., 2009). This further suggests that AMX can also be removed by charge neutralization.

4. Conclusions

This study proves that intermittent electrocoagulation is effective in the treatment of municipal wastewater contaminated with recalcitrant pharmaceuticals. Conventional pollutants namely organic compounds (COD and DOC), aromatic and humic substances (UV₂₅₄) and nutrients (NH₄-N and PO₄-P) were removed via adsorption onto aluminum coagulants, oxidation on the anode surface and precipitation. The treatment efficiencies of these pollutants improved with the current density and reaction



Figure 1. Removal efficiencies of a.) DCF, b.) CBZ and c.) AMX using continuous and intermittent (5 min ON/ 20 min OFF) electrocoagulation at varying current densities (Initial concentration of each pharmaceutical = 0.01 mg/L, HRT=3 hours)

time. Meanwhile, the removal of selected pharmaceuticals DCF, CBZ and AMX depends on their physicochemical characteristics. Hydrophobic CBZ has higher tendency to be adsorbed by the generated coagulants while AMX and DCF are more likely to be removed by charged neutralization and electro-flotation. Both continuous and intermittent DC application showed increasing removal as the current density increases. However, for practical reasons, intermittent electrocoagulation is more favorable as it minimizes energy consumption as well as the anode passivation phenomenon.

Acknowledgment

The authors would like to acknowledge the Sanitary and Environmental Engineering Division (SEED) of the Civil Engineering Department at UNISA for providing the facilities and research fund. This study was partially funded by the University of Salerno with the projects n. ORSA167105 and ORSA154525 and are linked to the project n. EG16MO01 funded by Italian Ministry of Foreign Affairs and International Cooperation. We also thank the University of the Philippines-Diliman and the Engineering Research and Development for Technology (ERDT) through the Department of Science and Technology-Philippines for the PhD Scholarship Grant and Sandwich Program awarded to BMB Ensano. The authors also deeply appreciate the technical support provided by dr. A. Farina, P. Napodano and M. Giordano all throughout the completion of this research.

References

- Acero J.L., Benitez F.J., Real F.J. and Teva F. (2016), Micropollutants removal from retentates generated in ultrafiltration and nanofiltration treatments of municipal secondary effluents by means of coagulation, oxidation, and adsorption processes, *Chemical Engineering Journal*, 289, 48–58.
- APAT, IRSA-CNR (2003). Metodi analitici per le acque. APAT, Rapporti 29/2003.
- Borea L., Naddeo V., Belgiorno V. (2017), Application of electrochemical processes to membrane bioreactors for improving nutrient removal and fouling control, *Environmental Science and Pollution Research*, 24, 321–333.
- Elabbas S., Ouazzani N., Mandi L., Berrekhis F., Perdicakis M., Pontvianne S., Pons M-N., Lapicque F., and Leclerc J-P. (2016), Treatment of highly concentrated tannery wastewater using electrocoagulation: Influence of the quality of aluminium used for the electrode, *Journal of Hazardous Materials*, 319, 69–77.
- Ensano B.M.B., Borea L., Naddeo V., Belgiorno V., de Luna M.D.G. and Ballesteros, F. Jr. (2016), Combination of Electrochemical Processes with Membrane Bioreactors for Wastewater Treatment and Fouling Control: A Review, *Frontiers in Environmental Science*, 4, 57.
- Ensano B.M.B., Borea L., Naddeo V., Belgiorno V., de Luna M.D.G. and Ballesteros, F. Jr. (2017), Removal of Pharmaceuticals from Wastewater by Intermittent Electrocoagulation, Water, 9, 85.
- Ganzenko O., Huguenot D., Hullebusch ED van., Esposito G., Oturan M. (2014), Electrochemical advanced oxidation and biological processes for wastewater treatment: a review of the combined approaches, *Environmental Science and Pollution Research*, 21, 8493–8524.

- Houtman C.J., Kroesbergen J., Lekkerkerker-Teunissen K., van der Hoek JP (2014), Human health risk assessment of the mixture of pharmaceuticals in Dutch drinking water and its sources based on frequent monitoring data, *Science of the Total Environment*, 496, 54–62.
- Hu D. and Wang L. (2016), Adsorption of amoxicillin onto quaternized cellulose from flax noil: Kinetic, equilibrium and thermodynamic study, *Journal of the Taiwan Institute of Chemical Engineers*, 64, 227–234.
- Inyang M., Flowers R., McAvoy D, Dickenson E. (2016), Biotransformation of trace organic compounds by activated sludge from a biological nutrient removal treatment system, *Bioresource Technology*, 216, 778–784
- Kobya M. and Demirbas E. (2015), Evaluations of operating parameters on treatment of can manufacturing wastewater by electrocoagulation, *Journal of Water Process Engineering*, 8, 64–74.
- Lin S.H., Wu C.L. (1996), Electrochemical removal of nitrite and ammonia for aquaculture, *Water Research*, 30, 715–721.
- Liu Y-J., Lo S-L., Liou Y-H., Hu C-Y. (2015), Removal of nonsteroidal anti-inflammatory drugs (NSAIDs) by electrocoagulation–flotation with a cationic surfactant, *Separation and Purification Technology*, 152, 148–154.
- Mohora E., Rončević S., Agbaba, J., Tubić A., Mitić M., Klašnja M. and Dalmacija B. (2014), Removal of arsenic from groundwater rich in natural organic matter (NOM) by continuous electrocoagulation/ flocculation (ECF), Separation and Purification Technology, 136, 150–156.
- Nghiem, L.D., Schäfer, A.I. and Elimelech, M. (2006) Role of electrostatic interactions in the retention of pharmaceutically active contaminants by a loose nanofiltration membrane, *Journal of Membrane Science*, 286, 52–59.
- Phan H.V., Hai F.I., McDonald J.A., Khan S.J., Zhang R., Price W.E., Broeckmann, A. and Nghiem, L.D. (2015), Nutrient and trace organic contaminant removal from wastewater of a resort town: Comparison between a pilot and a full scale membrane bioreactor. *International Biodeterioration & Bodegradation*, 102, 40–48.
- Pérez L. S., Rodriguez O.M., Reyna S., Sánchez-Salas J.L., Lozada D, Quiroz M.A. and Bandala E.R. (2015), Oil refinery wastewater treatment using coupled electrocoagulation and fixed films biological processes, *Physics and Chemistry of the Earth*, 91, 53-60.
- Rao Y., Yang H., Xue D., Guo Y., Qi F. and Ma J., (2016), Sonolytic and sonophotolytic degradation of Carbamazepine: Kinetic and mechanisms, *Ultrasonics Sonochemistry*, 32, 371–379.
- Putra E.K., Pranowo R., Sunarso J., Indraswati N. and Ismadji S. (2009), Performance of activated carbon and bentonite for adsorption of amoxicillin from wastewater: Mechanisms, isotherms and kinetics, *Water Research* 43, 2419–2430.
- Tafti A.D., Mirzaii S.M.S., Andalibi M.R. and Vossoughi M. (2015), Optimized coupling of an intermittent DC electric field with a membrane bioreactor for enhanced effluent quality and hindered membrane fouling, *Separation and Purification Technology*, 152, 7–13.
- Ulu F., Barışçı, S., Kobya, M., and Sillanpää, M. (2015), An evaluation on different origins of natural organic matters using various anodes by electrocoagulation, *Chemosphere*, 125, 108–114.